

Chemical speciation of iron in urban shallow groundwater, southeastern Brazzaville, Congo

Matini Laurent^{1,2}, Moutou Joseph Marie^{1,2}, Nkounkou Tomodiatounga Dominique³, and Massoumou Massolola¹

¹Laboratory of Applied Mineral Chemistry, Sciences and Techniques Faculty, Marien NGouabi University, P.O. Box 69 Brazzaville, Republic of the Congo

²Laboratory of chemistry, Superior National School, Marien NGouabi University, P.O. Box 69 Brazzaville, Republic of the Congo

³Mechanical, Energetic and Engineering laboratory, Higher National Polytechnic School, Marien NGouabi University, P.O. Box 69 Brazzaville, Republic of the Congo

Copyright © 2018 ISSR Journals. This is an open access article distributed under the *Creative Commons Attribution License*, which permits unrestricted use, distribution, and reproduction in any medium, provided the original work is properly cited.

ABSTRACT: This study assessed the groundwater quality in southeastern Brazzaville, Congo, in order to calculate iron chemical speciation. A total of twenty two groundwater samples were collected in rainy and dry seasons. Physico-chemical parameters namely pH, TDS, EC, T, D.O, TH, Alk, ORP, Ca²⁺, Mg²⁺, Na⁺, K⁺, Cl⁻, SO₄²⁻, HCO₃⁻, NO₃⁻, PO₄³⁻, F⁻, Fe_{tot}, Fe²⁺ and Fe³⁺ were analyzed. The results indicate acidic nature and oxidizing conditions in groundwater, contamination of groundwater by domestic sewage, water septic tanks at the sight of EC values and iron content higher than permissible limit of 0.3 mg/L in the majority of groundwater samples. Modeling of iron chemical speciation in the groundwater samples using Visual Minteq computer program revealed that the predominant species were free ionic form Fe²⁺ in both season followed by FeSO₄(aq) for Fe(II) species. The majority of Fe(III) species were fluoro complexes FeF²⁺ and (FeF₂)⁺, sulphato complex ion FeSO₄⁺ and free ionic form Fe³⁺ in rainy season. The hydroxo complexes of Fe(III) are predominant in dry season.

KEYWORDS: Groundwater, Iron, Chemical speciation, Modelling, Visual Minteq.

1 INTRODUCTION

Groundwater is one of natural water resource that human being use for domestic needs, industrial and agricultural activities. In developing countries many people use shallow groundwater sources for drinking and domestic purposes. The study of groundwater chemistry reveals much information on different purposes [1, 2]. Problems which affect the quality of groundwater are waste disposal on land surface, septic tanks, use of fertilizers and other agricultural chemicals, contamination by geogenic sources [3, 4]. The lithology of the area is one of the factors that control groundwater chemistry.

The impact of human activities has repercussions on the physical and chemical properties of the groundwater. Thus in the case of unsafe drinking water some diseases such as cholera, typhoid, dysentery and diarrhea can arise [5]. In developing country like Congo-Brazzaville, Soil waste dumps are frequently present and can affect the quality of groundwater through metal pollution. Iron is one of the metals found in groundwater at level high the permissible limit of 0.3 mg/L for drinking water in the country [6]. Iron is a micronutrient and the body requires it in small amounts for metabolic activities. When iron content in water is higher than 0.5 mg/L, severe toxicity in human physiological systems can occur [7]. Therefore it is primordial to determine groundwater quality in order to assess its suitability for a particular purpose [8].

Chemical speciation of metals in groundwater has been investigated in many studies using geochemical softwares such as Phreeqc, Vminteq, Wham [9-14]. In water systems metals can be coordinated with some anions. The chemical speciation of metals determines their toxicity and bioavailability which are a prediction of metal behavior and possible impact in the environment [15, 16].

labeled, transported to the laboratory and preserved in a refrigerator at a temperature of about 4 °C until analysis. All reagents used were of analytical grade. Samples were unfiltered and the concentration of iron could correspond to the total concentration if the groundwater was used by the consumers for drinking. The physico chemical parameters were determined using standard methods [18].

The measurements of pH, oxidation reduction potential (ORP), electrical conductivity (E.C), total dissolved solids (TDS), temperature (T) and dissolved oxygen (DO) were carried out in the field using Consort C933 multi-parameters portable. Total hardness (TH), Ca²⁺ was determined by EDTA complexometric titration and Mg²⁺ by calculation from total hardness and calcium. Na⁺ and K⁺ were determined by photometric method. Chloride (Cl⁻) in groundwater was determined by argentometric titration method using potassium chromate indicator solution and standard silver nitrate titrant, SO₄²⁻ by turbidimetric method. Bicarbonate (HCO₃⁻) was obtained by titration method after measured value of total alkalinity, methyl orange use as an indicator. Nitrate (NO₃⁻) and orthophosphates (PO₄³⁻) were determined by spectrophotometric method. Iron contents were determined by phenanthroline spectrophotometry method. The method involves the complexation of Fe²⁺ with 1, 10-phenanthroline to produce an intensely red-orange colored complex. Iron present in the water exists as Fe³⁺ and Fe²⁺. Before to measure total iron contents in groundwater samples, first Fe³⁺ was reduced to Fe²⁺ by the addition in excess of a reducing agent hydroxylamine to maintain iron in the +2 state. After that, ferric ions contents were calculated by subtraction. All the water quality parameters were expressed in mg/L, except EC (μS/cm) and pH (unit).

2.3 STATISTICAL ANALYSIS

Hierarchical clustering analysis (HCA) was used to get a better separation among groundwater samples. HCA is an exploratory technique with graphic output as a dendrogram and the method is used frequently in environmental sciences [19-23]. It helps to classify group samples or group variables of a data set. In this study, HCA was performed on the data set by means of the Ward's method, using Euclidean distances as a measure of similarity. The statistical software used is Statistica 7.1.

2.4 MODELING CALCULATIONS

The geochemical computer program Visual Minteq ver. 3.0 [12] was used to determinate the chemical speciation of iron that may exist in the groundwater. This computer program has an extensive database of aqueous species for interpretation purposes. The anion concentrations determined in the study influence the

3 RESULTS AND DISCUSSION

3.1 PHYSICO-CHEMICAL CHARACTERISTICS

Groundwater data of the samples in rainy and dry seasons were statistically described as minimum, maximum, mean, median, variance and standard deviation (SD) in Table 1 and Table 2, respectively. The results indicate that pH values varied from 2.87 to 5.83 with a mean of 4.11±0.91 in rainy season and from 4 to 6.73 with a mean of 5.08±0.86 in dry season. Groundwater is acidic in nature in both seasons with a more pronounced acidity in the rainy season. This could be explained by the decomposition of the soil organic matter (high vegetation in the study area) and a pyrite oxidation [25] as describe in the following reaction:



This equation describes pyrite oxidation in the presence of water leading to the formation of acid (H⁺) and sulfates (SO₄²⁻). ORP values of the groundwater samples ranged from 181.10 to 404.70 mV in rainy season and from 130.27 to 395 mV in dry season. These values characterize an oxidizing environment. The variable ORP contributed to the interaction between water and redox sensitive minerals such as sulphides. In rainy season 59.09% of groundwater samples have E.C value range from 305.5 to 544.6 μS/cm and in dry season 68.18% of the samples have E.C range from 302.67 to 704 μS/cm. Overall the high values of electrical conductivity (E.C > 300 μS/cm) suggests to contamination of groundwater by domestic sewage and water septic tanks. This could explain the high concentrations of sodium observed in the corresponding groundwater samples. TDS are mainly compounds of inorganic salts, the amount ranges from 34.57 to 289.67 mg/L and from 27.33 to 374.67 mg/L in rainy and dry seasons, respectively. With reference to quality, all the groundwater samples (100%) are freshwater with TDS values <1,000 mg/L [26]. Dissolved oxygen (DO) values in groundwater samples were small in rainy season as 0.03 to 0.32 mg/L and 3.2 to 8 mg/L in dry season, which denotes an intense bacterial activity in rainy season.

Total hardness (TH) ranged from 61.67 to 100.84 mg/L CaCO₃ in rainy season and from 74.92 to 330.52 mg/L in dry season. Water with hardness value above 300 mg/L may cause kidney and heart diseases [27]. According the classification of water depending upon hardness [28] as shown in Table 3, groundwater samples belong to soft (31.81% of samples) and medium hard (68.39 % of samples) in rainy season. In dry season, the classification was medium hard (40.90% of samples) and hard (59.10% of samples).

Alkalinity (Alk) is a measure of the ability of water to neutralize acids. The total alkalinity varied from 12.20 to 67.10 mg/L in rainy season and from 9.76 to 58.56 mg/L in dry season. These values denote a decreasing of the alkalinity from rainy season to dry season. In both seasons, the predominant cation trend in groundwater samples was Na⁺ > Ca²⁺ > Mg²⁺ > K⁺. Sodium (Na⁺) was the dominant cation with the concentration ranged between 48.56 to 88.44 mg/L with an average of 62.39 mg/L in rainy season and from 11.65 to 119.25 mg/L with an average of 50.98 mg/L in dry season. In all the groundwater samples, sodium concentration was lower than the maximum permissible limit of 200 mg/L in both seasons, according the WHO guideline. Source of Na⁺ and Cl⁻ in the groundwater of the study area may be domestic wastewater, which is usually enriched in Na⁺ relatively to Cl⁻. Calcium (Ca²⁺) was the second most dominant cation with concentration ranged from 7 to 18 mg/L with a mean of 12.59 mg/L in rainy season and from 10 to 54 mg/L with a mean of 26 mg/L in dry season. The other cations, magnesium (Mg²⁺) its concentration ranged from 7 to 17 mg/L with an average of 12.4 mg/L in rainy season and from 10 to 47 mg/L with an average of 22.64 mg/L in dry season. Potassium (K⁺) presents the lowest concentrations which ranged from 3.1 to 8.8 mg/L with an average of 5.6 mg/L in rainy season and from 1.9 to 8 mg/L with an average of 3.76 mg/L in dry season. Potassium has the tendency to be fixed by clay minerals which could explain the low concentrations [29]. All the concentrations of major cations are below the WHO limits [17].

Assessment of anion concentrations shows that bicarbonate (HCO₃⁻) was the dominant anion with concentration ranged from 10 to 55 mg/L with a mean of 23.59 mg/L in rainy season and from 8 to 48 mg/L with a mean of 27.50 in dry season. Chloride concentrations ranged from 3 to 21 mg/L with an average of 9.36 mg/L in rainy season and from 3 to 17 mg/L with an average of 9.82 mg/L in dry season. Sulphate concentration ranged from 4 to 13 mg/L with a mean of 9.09 mg/L in rainy season and from 0.32 to 16 mg/L with a mean of 5.72 mg/L in dry season. Sulphate concentrations in the studied area are attributed mainly to oxidation of pyrite (FeS) in the aquifer [25].

Table 1. Descriptives statistics of groundwater parameters in rainy season

Parameter	Min	Max	Mean	Median	Variance	SD
pH	2.87	5.83	4.11	3.98	0.82	0.91
TDS	34.57	289.67	157.95	164.5	5347.23	73.12
EC	67.25	544.67	298.55	310.67	19108.88	138.23
T	24.30	27.30	25.94	25.80	0.72	0.85
D.O	0.03	0.32	0.16	0.17	0.00	0.07
TH	61.67	100.84	82.05	82.09	128.77	11.35
Alk	12.20	67.10	28.78	25.62	192.03	13.86
ORP	181.10	404.70	321.20	343.90	3955.59	62.89
Ca ²⁺	7.00	18.00	12.59	12.00	9.59	3.10
Mg ²⁺	7.00	17.00	12.14	12.00	6.41	2.53
Na ⁺	48.56	88.44	62.39	61.15	85.26	9.23
K ⁺	3.10	8.80	5.60	5.50	2.75	1.66
Cl ⁻	3.00	21.00	9.36	7.50	31.19	5.59
SO ₄ ²⁻	4.00	13.00	9.09	10.00	6.66	2.58
HCO ₃ ⁻	10.00	55.00	23.59	21.00	129.02	11.36
NO ₃ ⁻	0.30	2.55	1.17	1.05	0.37	0.61
PO ₄ ³⁻	0.12	0.28	0.19	0.19	0.00	0.05
F ⁻	0.13	0.34	0.21	0.20	0.002	0.05
Fe _{tot}	0.03	0.32	0.16	0.17	0.005	0.07
Fe ²⁺	0.02	0.17	0.07	0.075	0.003	0.05
Fe ³⁺	0.01	0.15	0.09	0.095	0.002	0.02

Table 2. Descriptives statistics of groundwater parameters in dry season

Parameter	Min	Max	Mean	Median	Variance	SD
pH	4.00	6.73	5.08	4.81	0.75	0.86
TDS	27.33	374.67	173.05	169.83	7793.25	88.28
C.E	53.73	704.00	340.61	352.50	25647.62	160.15
T	22.90	25.30	24.00	24.05	0.42	0.654
D.O	3.20	8.00	6.00	6.35	1.75	1.32
TH	74.92	330.52	159.17	144.45	5590.78	74.77
Alk	9.76	58.56	33.55	35.38	178.01	13.34
ORP	130.27	395.00	286.61	305.35	5456.84	73.87
Ca ²⁺	10.00	54.00	26.00	23.00	203.33	14.26
Mg ²⁺	10.00	47.00	22.64	20.00	122.34	11.06
Na ⁺	11.65	119.25	50.98	53.22	933.69	30.56
K ⁺	1.90	8.00	3.76	3.25	2.04	1.43
Cl ⁻	3.00	17.00	9.82	10.00	14.63	3.83
SO ₄ ²⁻	0.32	16.00	5.72	5.00	25.38	5.04
HCO ₃ ⁻	8.00	48.00	27.50	29.00	119.60	10.94
NO ₃ ⁻	0.25	6.30	1.61	1.15	2.16	1.47
PO ₄ ³⁻	0.18	2.10	0.75	0.67	0.31	0.56
F ⁻	0.10	0.90	0.29	0.21	0.06	0.23
Fe _{tot}	0.15	0.90	0.31	0.20	0.06	0.25
Fe ²⁺	0.02	0.75	0.18	0.08	0.06	0.24
Fe ³⁺	0.13	0.15	0.13	0.12	0.00	0.01

Table 3. Classification of water according the hardness

Classification	Hardness range (mg/L CaCO ₃)
Soft	0 - 75
Medium hard	75 - 150
Hard	150 - 300
Very hard	> 300

In both seasons all the anion concentrations were below the WHO limits for drinking. In aerobic conditions, nitrates are the end product of the stabilization of organic nitrogen, NO₃⁻ concentration of groundwater samples varied from 0.3 to 2.55 mg/L and from 0.25 to 6.30 mg/L with a mean of 1.17 mg/L and 1.61 mg/L in rainy and dry seasons, respectively. In the studied area, nitrate concentration is below the WHO permissible limit of 50 mg/L. The orthophosphate (PO₄³⁻) concentration ranged from 0.12 to 0.28 mg/L and from 0.18 to 2.10 mg/L with a mean of 0.19 mg/L and 0.75 mg/L in rainy and dry seasons, respectively. Fluoride (F⁻) exhibited low concentration in groundwater samples with a range of 0.13 - 0.34 mg/L and an average of 0.2 mg/L in rainy season. In dry season the range of F⁻ concentration was 0.10 - 0.90 mg/L with an average of 0.29 mg/L.

The concentration of observed total iron (Fe_{tot}) in groundwater samples varied from 0.03 to 0.32 mg/L with a mean of 0.16 mg/L in rainy season while the range in dry season was 0.15 to 0.90 mg/L with a mean of 0.29 mg/L. Ionic iron (III) concentration as Fe³⁺ ranged from 0.01 to 0.15 mg/L with a mean of 0.09 mg/L in rainy season. In dry season, the range was 0.13 to 0.15 mg/L with a mean of 0.13 mg/L. In both seasons, the concentrations of major anions exhibited the following order: HCO₃⁻ > Cl⁻ > SO₄²⁻.

3.2 HIERARCHICAL CLUSTER ANALYSIS (HCA)

The spatial classes of groundwater samples are illustrated using a dendrogram. In rainy season, the dendrogram consists of three groups of very distinct groundwater samples (fig. 2). Group 1 is very characteristic of the other two because it contains the samples with the lowest alkalinity (fig. 3), lowest concentration of ammonium and fluoride ions (figs. 4 and 5) and the highest mineralization (EC, TDS) (fig. 6). The trend of mineralization was in the following decreasing order: group 1 > group 2 > group 3. These three clusters are also differentiated by the concentration of ammonium ion as: group 1 < group 2 < group 3. This suggests that degradation of organic matter is greater in water points of group 3 during the rainy season. The parameter Fe_{tot} discriminates also the three clusters.

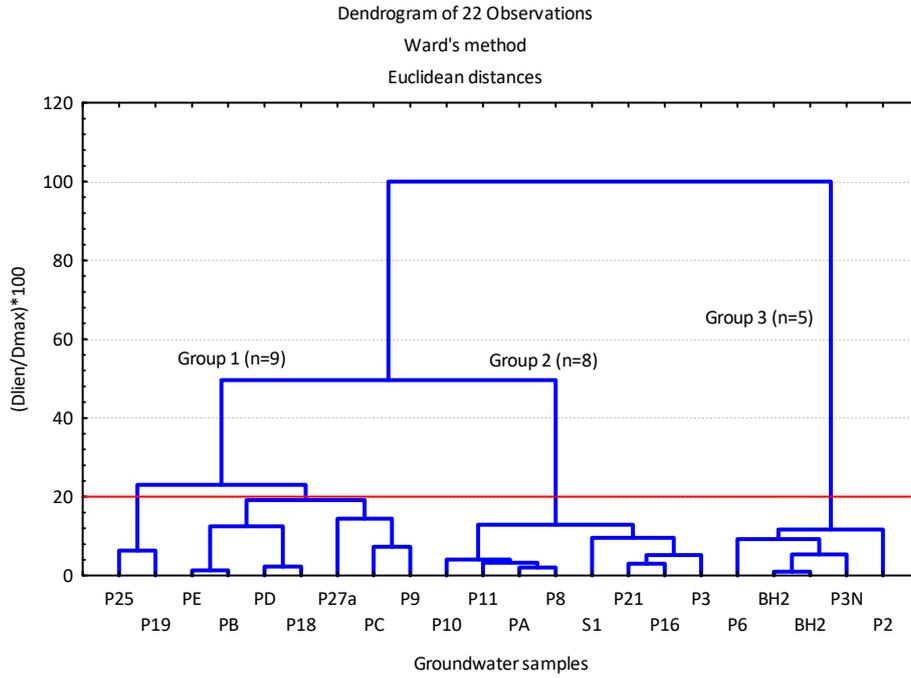


Fig. 2. Dendrogram of groundwater samples in rainy season

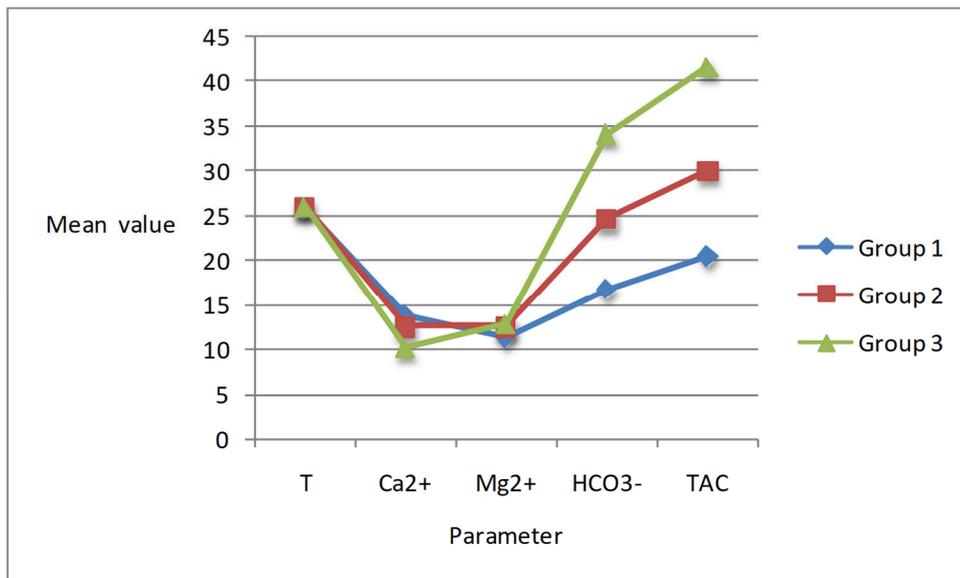


Fig. 3. Distribution of mean value of T, Ca²⁺, Mg²⁺, HCO₃⁻ and TAC in the three clusters (rainy season)

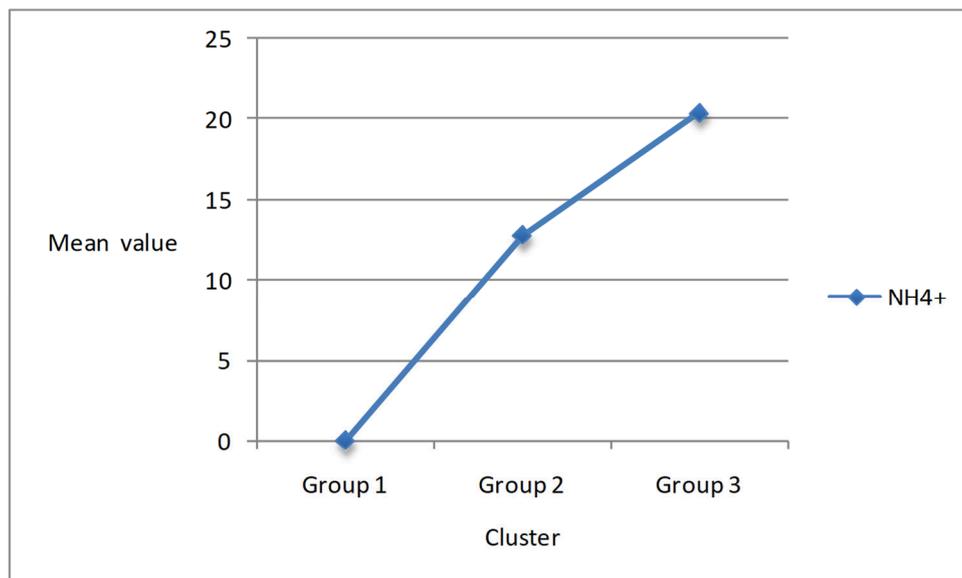


Fig. 4. Distribution of mean value of NH_4^+ concentration in the three clusters (rainy season)

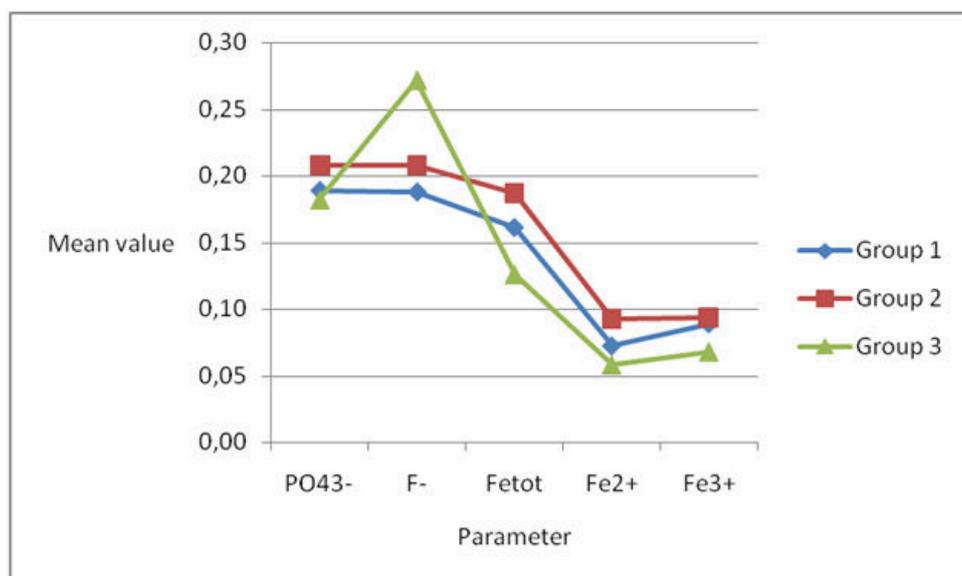


Fig. 5. Distribution of mean value of PO_4^{3-} , F^- , Fe_{tot} , Fe^{2+} and Fe^{3+} concentration in the three clusters (rainy season)

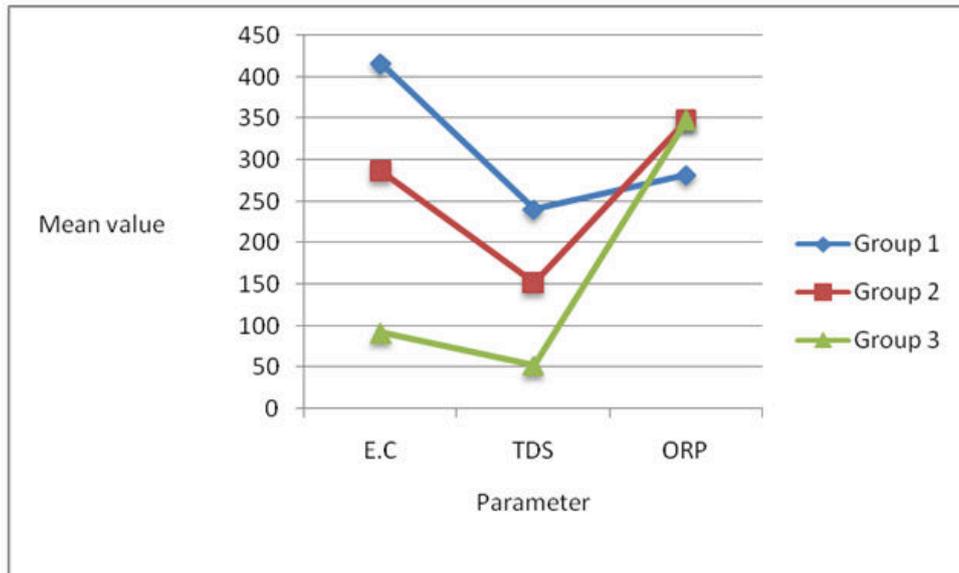


Fig. 6. Distribution of mean value of E.C, TDS and ORP in the three clusters (rainy season)

In dry season, the dendrogram presents four groups of groundwater samples (fig. 7). The parameters that discriminate the four groups of groundwater samples are major ions Ca^{2+} , Na^+ , Mg^{2+} (fig. 8) which translate the mineralization of water by the EC and TDS parameters and the hardness (TH) (fig. 9). This mineralization is in the order: group 1 > group 2 > group 3 > group 4. Water hardness also discriminates the four groups of groundwater samples whose the predominance is in the order: group 1 > group 2 > group 4 > group 3 (fig. 9). The classification of these groups of groundwater according to the hardness makes it possible to observe that the water samples in group 1 have an average hardness greater than 200 mg/L of CaCO_3 , which classifies them as hard water.

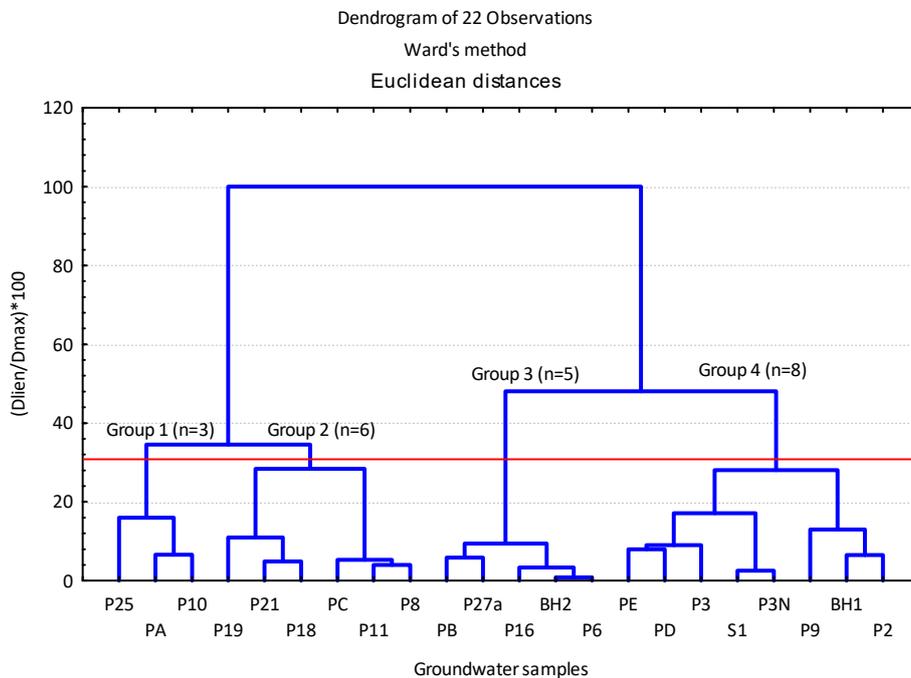


Fig. 7. Dendrogram of groundwater samples in dry season

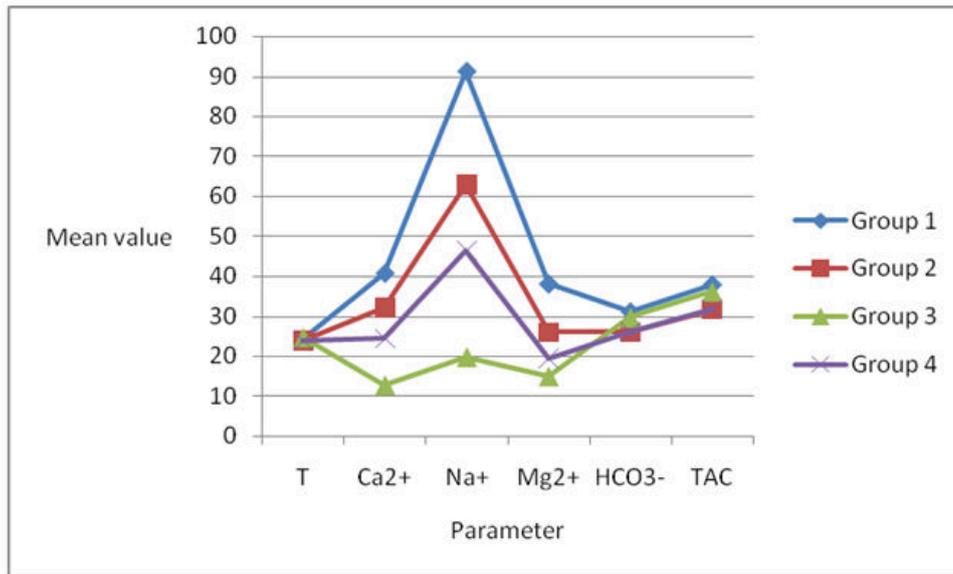


Fig. 8. Distribution of the mean value of T, Ca²⁺, Na⁺, Mg²⁺, HCO₃⁻ and TAC in the four clusters (dry season)

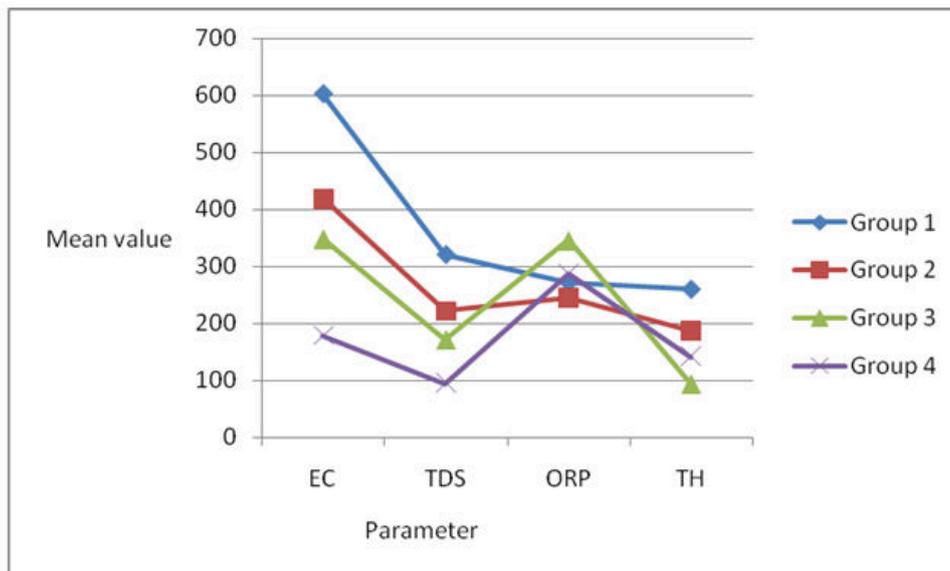


Fig. 9. Distribution of the mean value of EC, TDS, ORP and TH in the four clusters (dry season)

3.3 WATER TYPE

In the Schoeller diagram, taking into account the average composition of major ions, the facies of the groundwater samples in the different clusters are Na-HCO₃ and Na-Mg-HCO₃ in rainy season and dry season, respectively (figs. 10 and 11)

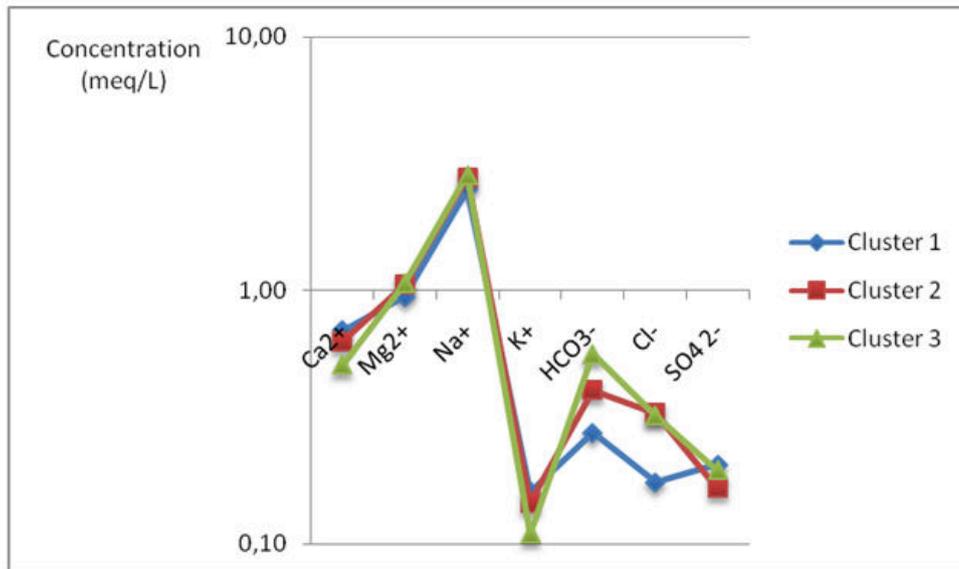


Fig. 10. Distribution of the average composition in major ions in the three clusters (rainy season)

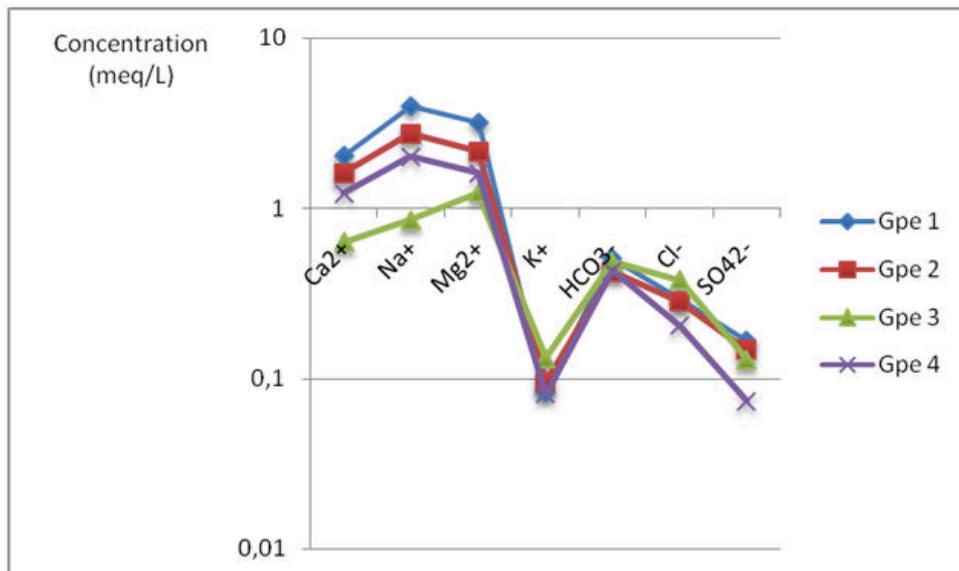


Fig. 11. Distribution of the average composition in major ions in the four clusters (dry season)

3.4 MODELING CHEMICAL SPECIATION OF IRON

The geochemical computer program Visual MINTEQ ver. 3.0 [12] was used to determinate the iron chemical species that may exist in the groundwater. This computer program has an extensive database of aqueous species for interpretation purposes. The input data take in account the parameters temperature, pH, and concentrations of Fe³⁺, Fe²⁺, Cl⁻, SO₄²⁻, F⁻ and PO₄³⁻.

In rainy season according the computations, Fe (II) was mainly present as: Fe²⁺ and FeSO₄ (aq) constituting 96.04% and 3.94%, respectively of the species. FeCl⁺ and FeF⁻ were present as traces (fig. 12). Under the pH range of 5 to 8, the predominant dissolved form of iron Fe(II) was observed in other studies as Fe²⁺ [30]. The predominant forms of Fe (III) with their percentage distribution were FeF²⁺ (40.89%), FeSO₄⁺ (36.49%), Fe³⁺ (15.21%) and FeF₂⁺ (7.15%). The forms FeF₃ (aq), FeCl⁺ and Fe(SO₄)₂⁻ were present with a percentage distribution less than 1% (fig. 13). Minor amounts of fluoride complexes were observed with trace constituents as iron in natural waters [31]. In water, fluoride acts as a strong anion and forms some soluble complexes with ferric ion (Fe³⁺) and pH is a dependant parameter in the formation of these soluble complexes [32].

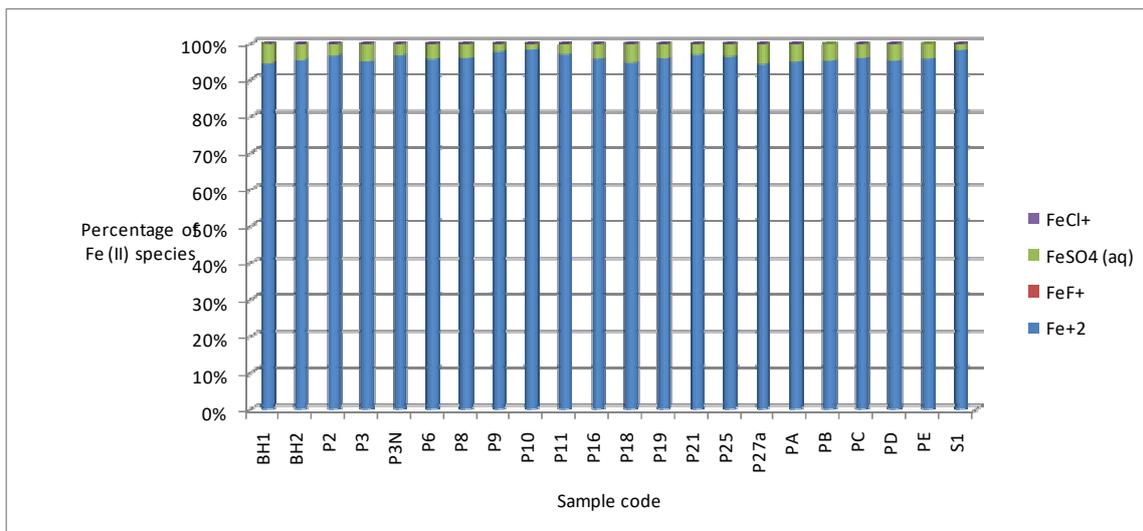


Fig. 12. Distribution of major Fe(II) species (rainy season)

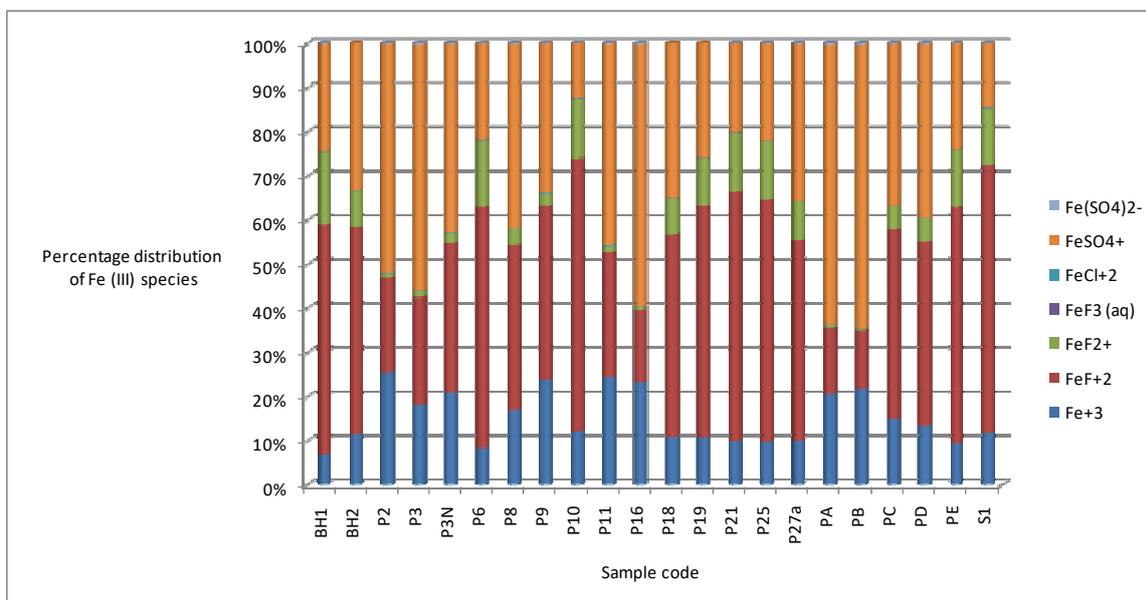


Fig. 13. Distribution of major Fe(III) species (rainy season)

In dry season, the majority of Fe(II) occurs as Fe^{2+} (98.47%), followed by melanterite $FeSO_4(aq)$ or $FeSO_4 \cdot 7H_2O$ (1.21%). The following forms: $FeCl^+$, FeF^+ , $FeOH^+$, $FeHPO_4(aq)$ and $FeH_2PO_4^+$ were present at trace levels in all the groundwater samples (fig. 14). The predominant species of Fe(III) were $Fe(OH)_2^+$ (83.35%), $FeOH^{2+}$ (11.86%) and $FeHPO_4^+$ (3.58%). Other forms such as $Fe(OH)_3(aq)$, FeF^{2+} , $(FeF_2)^+$, $FeF_3(aq)$, $FeCl^{2+}$, $FeSO_4^+$, $FeH_2PO_4^{2+}$, $FeCrO_4^+$ et $Fe(OH)_4^-$ occur at trace concentration (fig. 15). Thus Fe^{2+} is the most soluble and bioavailable form of Fe(II) in the pH range observed during both seasons. Majority of the Fe(III) species (95.21%) was found as hydroxo complexes.

Some studies have shown the predominance of ferrous ion and melanterite ($FeSO_4 \cdot 7H_2O$) in the calculation of chemical speciation of iron in groundwater [33], [34], [35], [36].

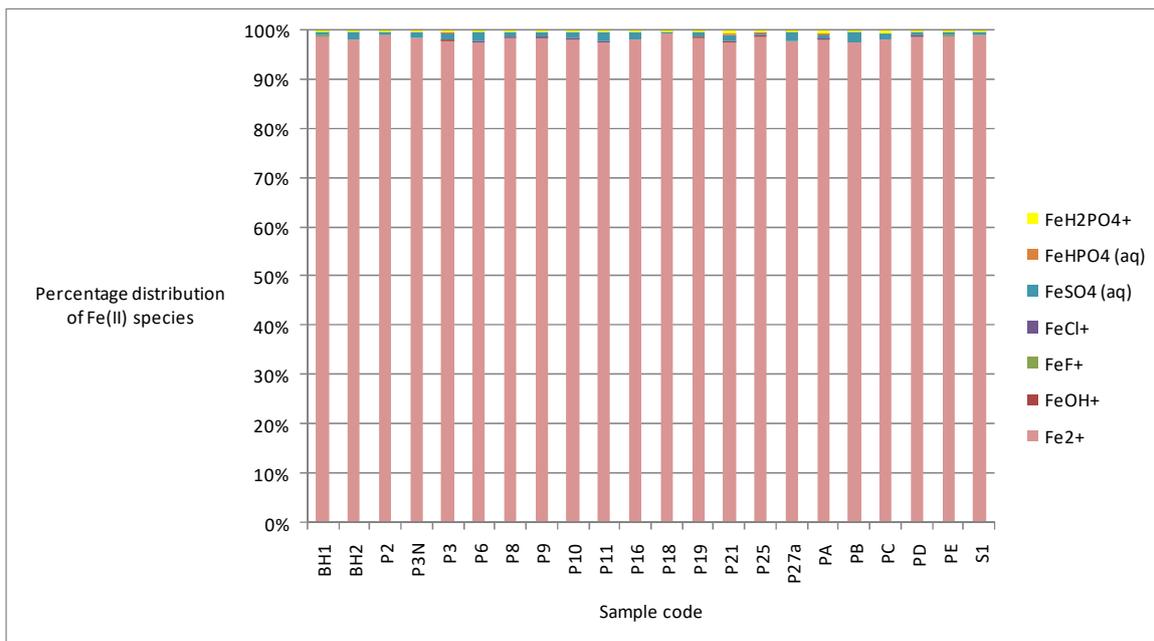


Fig. 14. Distribution of major Fe(II) species (dry season)

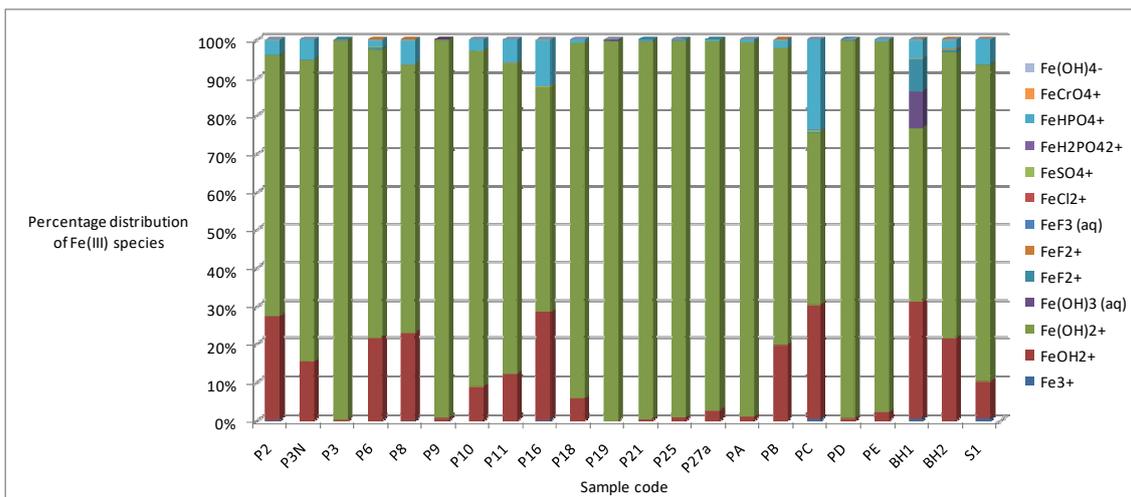


Fig. 15. Distribution of major Fe(III) species (dry season)

4 CONCLUSION

The hydrochemical analysis of the groundwater in the study area reveals that the groundwater is soft, more or less acidic tending towards neutrality in certain water points. The oxidizing conditions prevail in the groundwater during both seasons. Cluster analysis has highlighted the discriminant parameters among the groundwater sample groups, namely alkalinity, NH₄⁺ that reflect a decomposition of organic matter, fluoride and total iron in the rainy season. In dry season, the discriminant parameters between groups of groundwater samples were EC and TDS, which reflect mineralization, ORP and total hardness. Calculating the chemical speciation of iron in the groundwater samples using the Vminteq program indicate that the forms Fe²⁺ following by FeSO₄(aq) are predominant for Fe(II) in both seasons. The majority of Fe³⁺ ions in the groundwater samples is in complexed form. Fluoride complexes and the hydroxo complexes of Fe(III) occur mostly in the rainy season and are predominant in the dry season for Fe(III).

REFERENCES

- [1] V. Satish Kumar, B. Amarander, Ratnakar Dhakate, S. Sankara, K. Raj Kumar, " Assessment of groundwater quality for drinking and irrigation use in shallow hard rock aquifer of Pudunagaram, Palakkad District Kerala," *Applied Water Science*, vol. 6, pp. 149-167, 2016.
- [2] K. El oumlouki, R. Moussadek, A. Zouahri, H. Dakak, M. Chati, M. El amrani, " Étude de la qualité physico-chimique des eaux et des sols de la région Souss Massa, (Cas de périmètre Issen), Maroc," *Journal of Materials & Environmental Science*, vol. 5, no. S2. pp. 2365-2374, 2014.
- [3] J. J. Musa, J. J. Ahanonu, " Quality Assessment of Shallow Groundwater in Some Selected Agrarian Communities in Patigi Local Government Area, Nigeria," *International Journal of Basic and Applied Science*, vol. 1, no. 3, pp. 548-563, 2013.
- [4] Y. Gao, G. Yu, C. Luo, P. Zhou, " Groundwater Nitrogen Pollution and Assessment of Its Health Risks: A Case Study of a Typical Village in Rural-Urban Continuum, China," *PLoS One*, vol. 7, no. 4, pp. 1–8, 2012.
- [5] Muhammad Shoab, M. Javaid Asad, Sajid Aziz, Muhammad Usman, Abdur Rehman, Muhammad Mobeen Zafar and Muhammad Ilyas, " Prevalence of Pathogenic Microorganisms in Drinking Water of Rawalpindi and Islamabad," *World Journal of Fish and Marine Sciences*, vol. 8, no. 1, pp. 14-21, 2016.
- [6] WHO, Guidelines for drinking water quality, Recommendations, World Health Organization, 1984.
- [7] E. S. Gurzau, C. Neagu and A. E. Gurzau, " Essential metals-case study on iron," *Ecotoxicology and Environmental Safety*, vol. 6, no. 1, pp. 190-200, 2003.
- [8] M. Kumar Senthil and T. Rajkumar Jayavel, " Groundwater Quality Assessment in Lower Tamirabharani River Basin Tamil Nadu, India," *International Research Journal of Earth Sciences*, vol. 2, no. 10, pp. 8-15, 2014.
- [9] D. L. Parkhurst, C. A. J. Appelo, *User's guide to Phreeqc (version 2). A computer program for speciation, batch-reaction, one-dimensional transport, and inverse geochemical calculations*. U.S. Geological Survey. Water-Resources Investigations Report 99-4259, 1999.
- [10] B. A. Butler, J. F. Ranville and P. E. Ross, " Observed And Modeled Seasonal Trends In Dissolved And Particulate Cu, Fe, Mn, And Zn In A Mining-Impacted Stream," *Water Research*, vol. 42, no. 12, pp. 3135–3145, 2008.
- [11] Gabriel Ike Obiefuna and Donatus Maduka Orzulike, " Chemical Speciation of Some metal ions in Groundwaters of Yola Area Using Geochemical Model," *Journal of Applied Sciences and Environmental Management*, vol. 14, no. 2, pp. 65-70, 2010.
- [12] J. P. Gustafson, *Visual MINTEQ ver. 3.0.*, 2011. [Online] Available: <http://www2.lwr.kth.se/English/OurSoftware/vminteq> (February 10, 2018)
- [13] A. S. Ekwere and A. Edet, " Distribution and Chemical Speciation of Some Elements in the Ground Waters of Oban Area (South-Eastern Nigeria)," *Research Journal of Environmental and Earth Sciences*, vol. 4, no. 3, pp. 207-214, 2012.
- [14] I. U. Umoren and P. C. Onianwa, " Surface water quality status and chemical speciation of Qua Iboe River system, Niger-Delta, Nigeria," *Elixir Pollution*, vol. 47, pp. 9064-9073, 2012.
- [15] U. Benson Nsikak, U. Winifred Anake and O. Ifedolapo Olanrewaju, " Analytical Relevance of Trace Metal Speciation in Environmental and Biophysicochemical Systems," *American Journal of Analytical Chemistry*, vol. 4, pp. 633-641, 2013.
- [16] G. O. Adewuyi and M. T. Osobamiro, " Chemical Speciation and Potential Mobility of Some Toxic Metals in Tropical Agricultural Soil," *Research Journal of Environmental Toxicology*, vol. 10, pp. 159-165, 2016.
- [17] World Health Organization, *Guidelines for drinking-water quality : incorporating 1st and 2nd addenda, Vol. 1, recommendations*, 3rd ed. World Health Organization, 2008.
- [18] APHA, AWWA, WEF, *Standard methods for the examination of water and wastewater*, 19th ed. Washington, 1995.
- [19] N. A. Al-Odaini, M. P. Zakaria, M. A. Zali, H. Juahir, M. I. Yaziz and S. Surif, " Application of chemometrics in understanding the spatial distribution of human pharmaceuticals in surface water," *Environmental Monitoring and Assessment*, vol. 184, no. 11, pp. 6735-6748, 2012.
- [20] L. Yang, X. U. Linyu and L. Shun, " Water Quality Analysis of the Songhua River Basin Using Multivariate Techniques," *Journal of Water Resource and Protection*, vol. 1, no. 2, pp. 110-121, 2009.
- [21] E. G. Farmaki, N. S. Thomaidis, V. Simeonov and C. E. Efstathiou, " A comparative chemometric study for water quality expertise of the Athenian water reservoirs," *Environmental Monitoring and Assessment*, vol. 184, no. 12, pp. 7635-7652, 2012.
- [22] A. Z. Aris, S. M. Praveena, N. M. Isa, W. Y. Lim, H. Juahir, M. K. Yusoff and A. Mustapha, " Application of environmetric methods to surface water quality assessment of Langkawi Geopark (Malaysia)," *Environmental Forensics*, vol. 14, no. 3, pp. 230–239, 2013.
- [23] S. Pati, M. K. Dash, C. K. Mukherjee, B. Dash and S. Pokhrel, *Environmental Monitoring and Assessment*, vol. 186, no. 10, pp. 6385-6402, 2014.
- [24] S. Tepavitcharova, T. Todorov, M. Dassenakis and V. Paraskevopoulou, " Chemical speciation in waters influenced by lead-zinc metallurgical industry," *Environmental Monitoring and Assessment*, vol. 169, no. 1-4, pp. 27-36, 2010.

- [25] Egon Campos Dos Santos, Juliana Cecelia de Mendonca Silva and Hélio Anderson Duarte, " Pyrite Oxidation Mechanism by Oxygen in Aqueous Medium," *Journal of Physical Chemistry C*, vol. 120, pp. 2760–2768, 2016.
- [26] R. A. Freeze, J. A. Cherry, *Groundwater*, 2nd Ed. Prentice Hall Inc., Englewood, NJ., USA., 1979.
- [27] K. Ramesh, L. Elango, " Drinking water quality and chronic kidney disease of unknown etiology (CKDu): synergic effects of fluoride, cadmium and hardness of water," *International Journal of Environment and Pollution*, vol. 26, no. 6, pp. 497-504, 2006.
- [28] WHO, *Guideline for Drinking Water Quality*. 4th Ed. Geneva: World Health Organization, 2011.
- [29] G. Buyuk, J. Ryan, H. Ibrikci, K. Takashi, N. Güzel and E. Akca, " Extractable potassium and its relation to clays of Mesaoria plain soils, Cyprus," *INTERNATIONAL JOURNAL OF AGRICULTURE & BIOLOGY*, vol. 12, pp. 435–438, 2010.
- [30] I.M.I. Nagwa, " The Relations Between Concentration of Iron and the pH Ground Water (Case Study Zulfi Ground Water)," *International Journal of Environmental Monitoring and Analysis*, vol. 4, no. 6, pp. 140-145, .2016.
- [31] M. Edmunds and P. Smedley, *Fluoride in Natural Waters Occurrence, Controls and Health Aspects*, In: O. Selinus, B. Alloway, J. A. Centeno, R. B. Finkelman, R. Fuge, U. Lindh and O. Smedley (Eds.), *Essentials of Medical Geology*, Elsevier, Amsterdam: Academic Press, pp. 301-329, 2005.
- [32] A. Criaud and C. Fouillac, " Etude des eaux thermominérales carbogazeuses du Massif Central Français. I. Potentiel d'oxydo-réduction et comportement du fer," *Geochimica et Cosmochimica Acta*, vol. 50, no. 4, pp. 525-533, 1986.
- [33] J. Majzlan, C. B. Satish Mynemi, " Speciation of iron and sulfate in acid waters: Aqueous clusters to mineral precipitates," *Environmental Science & Technology*, vol. 39, pp. 188-194, 2005.
- [34] G. Christopher Hubbard, Stuart Black, L. Max Coleman, " Aqueous geochemistry and oxygen isotope compositions of acid mine drainage from the Río Tinto, SW Spain, highlight inconsistencies in current models," *Chemical Geology*. Vol. 265, nos. 3-4, pp. 321–334, 2009.
- [35] A.E. Edet, B.J. Merkel, O.E. Offiong, " Contamination risk assessment of fresh groundwater using the distribution and chemical speciation of some potentially toxic elements in Calabar (southern Nigeria)," *Environmental Geology*, vol. 45, pp. 1025–1035, 2004.
- [36] R. B. Jr. Herbert, " Seasonal variations in the composition of mine drainage-contaminated groundwater in Dalarna, Sweden," *Journal of Geochemical Exploration*, vol. 90, no. 3, pp. 197-214, 2006.